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# In Situ Measurement of Dissolved Methane and Carbon Dioxide in Freshwater Ecosystems by Off-Axis Integrated Cavity Output Spectroscopy

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**Supporting Information** 

**ABSTRACT:** A novel low-cost method for the combined, real-time, and in situ determination of dissolved methane and carbon dioxide concentrations in freshwater ecosystems was designed and developed. This method is based on the continuous sampling of water from a freshwater ecosystem to a gas/liquid exchange membrane. Dissolved gas is transferred through the membrane to a continuous flow of high purity nitrogen, which is then measured by an off-axis integrated cavity output spectrometer (OA-ICOS). This method, called M-ICOS, was carefully tested in a laboratory and was subsequently applied to four lakes in Mexico and



Alaska with contrasting climates, ecologies, and morphologies. The M-ICOS method allowed for the determination of dissolved methane and carbon dioxide concentrations with a frequency of 1 Hz and with a method detection limit of  $2.76 \times 10^{-10}$  mol L<sup>-1</sup> for methane and  $1.5 \times 10^{-7}$  mol L<sup>-1</sup> for carbon dioxide. These detection limits are below saturated concentrations with respect to the atmosphere and significantly lower than the minimum concentrations previously reported in lakes. The method is easily operable by a single person from a small boat, and the small size of the suction probe allows the determination of dissolved gases with a minimized impact on shallow freshwater ecosystems.

# 1. INTRODUCTION

Methane (CH<sub>4</sub>) is a potent greenhouse gas that contributes about 20% of the warming induced by greenhouse gases.<sup>1</sup> An important fraction of CH<sub>4</sub> emissions comes from natural sources, and it has been estimated that natural ecosystems emit approximately 160 Tg CH<sub>4</sub> yr<sup>-1.<sup>2</sup></sup>. Within these levels, it is estimated that lakes and reservoirs emit about 92 Tg CH<sub>4</sub> y<sup>-1.<sup>3</sup></sup>.

CH<sub>4</sub> emissions from lakes and reservoirs depend on numerous processes involved in biogeochemical carbon cycling. For instance, the balance of CH<sub>4</sub> production by methanogens vs CH<sub>4</sub> oxidation by methanotrophs, two major counteractive processes, strongly control dissolved CH<sub>4</sub> concentrations in lake water.<sup>4,5</sup> Quantification of the resulting dissolved CH<sub>4</sub> concentration ( $C_{CH4}$ ) throughout the water column is an important step in understanding the complexity of CH<sub>4</sub> cycling in freshwater ecosystems. Quantification of  $C_{CH4}$  ultimately allows the quantification of total diffusive CH<sub>4</sub> emissions to the atmosphere<sup>6,7</sup> or can be used as a pollution indicator.<sup>8</sup> Overall,  $C_{CH4}$  in lakes ranges usually from 1.00 × 10<sup>-8</sup> to 3.00 × 10<sup>-3</sup> mol L<sup>-1.9</sup> This is more than  $C_{CH4}$  in equilibrium with atmospheric CH4, which varies between 2.6  $\times$  10  $^{-9}$  and 4.00  $\times$  10  $^{-9}$  mol L  $^{-1.10-12}$ 

The determination of dissolved  $CO_2$  ( $C_{CO2}$ ) is equally important, even if  $CO_2$  emissions from freshwater ecosystems are often low compared to  $CH_4$  emissions.<sup>13</sup>  $CO_2$  is a central molecule of the carbon cycle, since it is the product of most biogeochemical processes, both aerobic and anaerobic, and the carbon source of several autotrophic processes, including primary production. Additionally,  $C_{CO2}$ , combined with other parameters, gives valuable information about bioprocesses occurring in an ecosystem. This is, for instance, the case of the respiratory quotient  $(Rq)^{14}$  and of the ratio between  $CO_2$ and  $CH_4$  concentrations,<sup>15</sup> both being indicators of aerobic/ anaerobic processes.

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Several methods have been used to measure  $C_{CH4}$  and  $C_{CO2}$  since the early 1960s.<sup>16</sup> The main methods that focus on dissolved CH4 have been listed in detail.<sup>17</sup> Most of these methods are based on water/gas equilibration, followed by gas phase measurement, since gases are often poorly soluble in water; therefore, the water/gas equilibrium favors the latter. Additionally, most gases are easier to detect in the gas phase than dissolved in liquids. After the dissolved gas has been transferred to the gas phase, several detection methods have been used. These include standard gas chromatography,<sup>18</sup> mass spectrometry,<sup>19</sup> and laser based detectors, including tunable diode laser absorption spectroscopy.9 Recently, significant advances have been made in the field of in situ quantification of dissolved gases, using underwater mass spectrometers<sup>20</sup> and off axis integrated cavity output spectrometers (OA-ICOS).<sup>21</sup> With the latter, the authors reported the novel application of an underwater OA-ICOS detector for the in situ measurement of dissolved CH<sub>4</sub> concentrations and isotopic compositions in the deep ocean. This method is now commercially available (Los Gatos Research Inc., Mountain View, CA).

These in situ methods represent a major breakthrough in the field and have a strong potential for application in challenging environments, such as the deep ocean. However, these methods are relatively expensive, require a remotely operated vehicle or a large frame, and have a response time of several minutes. In addition, at present, spectroscopic data analysis is performed off-line, not in real time, after the data is transferred to an external computer. Their application to freshwater ecosystems, which present a unique set of challenges, seems difficult in their current configuration. Freshwater ecosystems are often shallow; relatively small; located in remote locations; and morphologically complex, uneven, and variable relative to deep water bodies.<sup>22</sup> In such environments, a low-cost, lightweight, lowpower, fast-response and simple-to-use detector, operable from a small boat, and based on a miniature probe to avoid environmental perturbation would represent a significant improvement to the previously reported methods. Furthermore, as freshwater ecosystems are subject to high spatial and temporal variations,<sup>23,24</sup> a high data acquisition rate, short response time, real-time data reporting, and wide dynamic range are additional requirements for correct and detailed appraisals in these diverse and complex ecosystems.

The objective of the present work is to develop and deploy a low-cost in situ detector for the combined measurement of  $C_{CH4}$  and  $C_{CO2}$ , based on an ultraportable OA-ICOS analyzer. As shown hereafter, this method employs a membrane gas/ liquid exchange module. It was developed and tested in the laboratory, before field-testing in subtropical (Mexico) and boreal and tundra lakes (Alaska) with contrasting climates, ecologies, and morphologies. We demonstrate high data acquisition frequency and the precision and accuracy of the method.

#### 2. MATERIALS AND METHODS

**2.1. Detector and Prototype.** We used an OA-ICOS ultraportable greenhouse gas analyzer (UGGA, model 915-0011, Los Gatos Research Inc.A) to detect and quantify  $CH_4$ ,  $CO_2$ , and water vapor. This analyzer is ultraportable (essentially crushproof package and 15 kg weight), battery operated (70 W), and includes an internal vacuum pump for gas sampling, with a characteristic response time (time required to reach steady-state readings) of approximately 8 s. Details on the UGGA are presented in the Supporting Information (SI).

The UGGA was remotely operated and controlled from a tablet computer. The prototype was based on the transfer of the dissolved gases from a water sample to a gas phase, followed by the analysis of the gas phase by the UGGA. With this purpose, the prototype included a continuous flow of  $CH_4$ - and  $CO_2$ -free analytical grade nitrogen (Infra, Mexico or Airgas), controlled by a mass flow controller (GFC17, Aalborg) and a continuous flow of water extracted at the desired depth from the freshwater ecosystems through a vacuum line (Figure 1).



**Figure 1.** Prototype for dissolved  $CH_4$  and  $CO_2$  concentration measurements: 1.  $CH_4$  free nitrogen; 2. pressure control; 3. mass flow controller; 4. septum port for gas sampling/injection; 5. membrane filter; 6. vacuum control; 7. additional membrane filter; 8. gas/liquid exchange module; 9. water filter; 10. water sampling port; 11. disposable syringe for water sampling and headspace injection; 12. temperature measurement; 13. liquid recollection tank with volume control; 14. portable vacuum pump; 15. tablet remote operation of the detector; 16. Ultraportable greenhouse gas analyzer; A, B, C, D. flow control 3-way valves.

The gas and the liquid crossed at a gas exchange station that will be described below. The liquid flow rate was controlled through a volumetric flask and by the measurement of the time required for the water sample to reach the gas exchange station. Details on the prototype are presented in the SI.

To allow for easy field calibration, the prototype was used with two different modes of operation. The first, headspace equilibration combined with ICOS (H-ICOS), was a discrete sample measurement method, adapted from the traditional gas/ liquid equilibration technique. While the water line was continuously operating, a 60 mL water sample from the desired depth was taken with a 60 mL disposable syringe from the water sampling port no. 10 (Figure 1). The sample was evacuated and replaced by a fresh sample. Then, 20 mL of the liquid content of the syringe was evacuated and replaced by CH<sub>4</sub>- and CO<sub>2</sub>-free nitrogen, taken from septum port no. 4. The gas and liquid volumes were recorded, and the syringe was vigorously shaken for 20 s to allow for gas/liquid equilibration. Then, 15 mL of the 20 mL headspace of the syringe was injected in the gas line through port no. 4. The injection of that sample in the gas line caused a peak response (in ppm) of the UGGA that was integrated to determine the headspace CH<sub>4</sub> and CO<sub>2</sub> concentrations, in a way similar to, for instance, standard gas chromatography methods. After injecting the headspace sample in the UGGA, the temperature of the water sample in the syringe was determined, and the dissolved gas concentration in the original water sample was determined according to Henry's law (SI Equations S1 and S2).

The second method, membrane combined with ICOS (M-ICOS), was a continuous measurement method that consisted of a counter flow of  $CH_{4^-}$  and  $CO_2$ -free nitrogen and water continuously extracted by a vacuum pump at the desired depth,

crossing in a Permselect module (PDMSXA-1000, Medarray Inc.; Figure 1, no. 8). This exchange module was composed of an array of approximately 1250 silicone hollow fibers of 190  $\mu$ m internal diameter, 55  $\mu$ m thickness, with a total exchange area of 1000 cm<sup>2</sup>. The water flowed outside of the hollow fibers, inside the module's shell, while the CH<sub>4</sub>- and CO<sub>2</sub>-free nitrogen flowed inside the hollow fibers. Because of diffusive forces, the dissolved CH<sub>4</sub> and CO<sub>2</sub> contained in the water were transferred to the gas phase, where they were detected by the UGGA.

The gas transfer can be described by a diffusion model according to Fick's second law (eq 1), where dM/dt is the mass transfer rate (mol s<sup>-1</sup>); 1000 is a unit conversion factor from mol L<sup>-1</sup> to mol m<sup>-3</sup>; *K* is the membrane transfer coefficient (m s<sup>-1</sup>);  $A_M$  is the area of the membrane (m<sup>2</sup>);  $C_w$  is the dissolved gas concentration in the water sample (CH<sub>4</sub> or CO<sub>2</sub>; mol L<sup>-1</sup>);  $C_g$  is the gas concentration in the gas phase (mol L<sup>-1</sup>); and *H'* is the CH<sub>4</sub> and CO<sub>2</sub> air/water partition coefficient (–), defined from SI eq S2.

$$\frac{\mathrm{d}M}{\mathrm{d}t} = 1000 \cdot K \cdot A_{\mathrm{M}} \cdot \left( C_{\mathrm{w}} - \frac{C_{\mathrm{g}}}{H'} \right) \tag{1}$$

As shown in the SI, a direct proportionality between  $C_w$  and  $C_g$  can be established from eq 1, giving eq 2, where  $\alpha$  (dimensionless) is the proportionality parameter combining all membrane, gas, and water transfer characteristics ( $Q_g$ , K,  $A_M$  and H'), for an easier calculation.

$$C_{\rm w} = C_{\rm g} \cdot \alpha \tag{2}$$

The parameter  $\alpha$  can be determined by measuring first  $C_w$  by the H-ICOS method and then, with the same sample, by measuring  $C_g$  by the M-ICOS. This can be easily done in the field with actual water samples.

The M-ICOS method is subject to a delay time  $(t_d)$  between the time that the sample is actually extracted and the time that it reaches the gas/liquid module, where it is measured. Furthermore, there is an additional response time corresponding to the time required to reach steady-state readings at the UGGA. As will be shown in the Results section, a continuous flow stirred tank reactor (CSTR) model of hydraulic residence time<sup>25</sup> describes the hydraulic behavior of the system (eq 3) well, where  $C_{wm}$  is the dissolved gas concentration measured in the water (mol L<sup>-1</sup>) and  $t_r$  is the response time of the system, which can be also taken as the hydraulic residence time of the prototype.

$$C_{\rm wm} = C_{\rm w} \cdot \left[ 1 - \exp\left(-\frac{t}{t_{\rm r}}\right) \right]$$
(3)

After derivation of eq 3, eq 4 shows that online and real-time measurement of  $C_w$  can be obtained, thus avoiding long delay times between samples (see SI for details).

$$C_{\mathrm{w,t}} = \frac{\mathrm{d}C_{\mathrm{wm,t+t_d}}}{\mathrm{d}t} \cdot t_r + C_{\mathrm{wm,t+t_d}} \tag{4}$$

**2.2. Laboratory Testing.** The laboratory testing of both H-ICOS and M-ICOS methods is described in detail in the SI. Briefly, we tested the precision and linearity of the UGGA by injecting several  $CH_4$  and  $CO_2$  standards and by determining the signal-to-noise ratio at several gas concentrations. We then tested the H-ICOS concept by establishing the peak response of the UGGA to several volumes and  $CH_4$  and  $CO_2$  concentrations injected in the gas line. Next, synthetic water

samples with a known  $C_w$  were prepared in a lab-scale stirred tank reactor (STR) by injecting a continuous flow of standard gases in tap water with strong mixing (800 rpm), until saturation was reached. C<sub>w</sub> in these water samples was theoretically established according to Henry's Law (SI eqs S1 and S2). With these water samples, we established the time necessary to reach equilibrium between the water sample and the headspace of the sampling syringe, which is a basic requirement of the H-ICOS method. We also tested the H-ICOS method by comparing the measured  $C_w$  to the theoretical concentrations. We used the same samples to test the M-ICOS method. Measurements completed with the M-ICOS method were also compared to the theoretical concentrations, and the parameter  $\alpha$  was determined. The  $t_d$  and  $t_r$  of the M-ICOS method were established by switching between water containing CH<sub>4</sub> and CO<sub>2</sub>, and degassed water using a 3-way valve. These experiments were used to check the developed  $t_r$ model (eq 4).

2.3. Field-Testing. In order to validate the method in real case scenarios and to provide a demonstration of how the instrument operates under a range of field conditions, the prototype and both methods were tested in four lakes with contrasting climates, ecologies, and morphologies: (i) a eutrophic subtropical reservoir located in the Mexico metropolitan area (Lake Guadalupe); (ii) a mesotrophic subtropical reservoir located in the same drainage basin as Lake Guadalupe (Lake Llano); (iii) a shallow Alaskan thermokarst lake located in the boreal zone (Lake Goldstream); and (iv) a shallow nonthermokarst Alaskan lake located in the tundra (Lake Otto). Field studies were done in July 2013 (Lake Guadalupe and Lake Llano) and in August 2013 (Lake Otto and Lake Goldstream). In all lakes, C<sub>w</sub> profiles were determined by the M-ICOS method after determination of  $\alpha$  with the H-ICOS method (see Results and Discussion section). The profile procedure that best worked was as follows: the probe was maintained a few centimeters below the water surface for about 30 s; then the probe was lowered slowly and steadily by hand to the bottom of the lake, where it was maintained for an additional 30 s. A controlled diving speed was maintained, and to know the approximate depth corresponding to each concentration data, the time at each 0.5 m intermediary depths was noted. The diving speed was about 0.6 m min<sup>-1</sup>. With this procedure, about 100 data were acquired for each m of water column depth. The  $C_w$  data were corrected according to eq 4 before being interpreted.

We also measured in each lake the dissolved oxygen (DO) and pH profiles with a multiparametric probe (YSI 556 MP5, YSI, Yellow Springs, OH, in Mexico or Hydrolab Data Sonde, Hach Hydromet, Loveland, CO, in Alaska). Rq, expressed as the ratio between  $CO_2$  production and  $O_2$  consumption (mol mol<sup>-1</sup>)<sup>26</sup> was determined according to eq 5,<sup>15</sup> where  $C*_{CO2}$  and DO\* are the  $CO_2$  and DO concentration in equilibrium with the atmosphere.

$$Rq = \frac{C_{CO2} - C_{CO2}^*}{DO^* - DO}$$
(5)

#### 3. RESULTS AND DISCUSSION

**3.1. Laboratory Testing.** The injection of several standard gases from 2 to 500 ppm of  $CH_4$  and 20 to 1500 ppm of  $CO_2$  in the UGGA showed that the latter did not require further calibration, apart from its original factory calibration. The

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UGGA gave a linear response in the entire range tested. The signal-to-noise ratio, measured over 10 min for all standard gases, was  $1520 \pm 415$  for CH<sub>4</sub> and  $1803 \pm 344$  for CO<sub>2</sub>. We tested the peak response of the UGGA to the injection of several CH<sub>4</sub> and CO<sub>2</sub> guantities, which is the core concept of the H-ICOS method. SI Figure S1A shows an example of the UGGA response to the triplicate injection of 5 mL nitrogen, containing 2 ppm of CH<sub>4</sub>, which corresponds to an injected quantity of  $4.31 \times 10^{-10}$  mol. SI Figure S1B indicates the linear response of the UGGA to a range of  $8.31 \times 10^{-12}$  to  $1.63 \times$  $10^{-8}$  mol CH<sub>4</sub> and to a range of 2.61 ×  $10^{-9}$  to  $1.33 \times 10^{-6}$  mol CO2. The linear response of the UGGA for both gases indicated the validity of CH4 and CO2 quantification by peak injections. With these results, we estimated the method detection limit (MDL),<sup>27</sup> of the H-ICOS method. The minimum CH<sub>4</sub> quantity that was distinguishable from background noise with 99% confidence was  $7.63 \times 10^{-12}$  mol. An example of UGGA response to that CH<sub>4</sub> quantity is shown in inner SI Figure S1A. By using eqs 1 and 2 and under the experimental conditions, this MDL corresponds to a C<sub>CH4</sub> of  $2.69 \times 10^{-10}$  mol L<sup>-1</sup>. This MDL is significantly lower than the range of C<sub>CH4</sub> reported in lakes<sup>9</sup> and about 35 times less than the minimum  $C_{CH4}$  reported in takes and about 35 times less than the minimum  $C_{CH4}$  concentration of  $1.00 \times 10^{-8}$  mol L<sup>-1</sup> among the lowest  $C_{CH4}$  reported.<sup>11,28,29</sup> This detection limit is also significantly lower than the  $C_{CH4}$  at equilibrium with atmospheric CH<sub>4</sub>, which is  $2.80 \times 10^{-9}$  mol L<sup>-1</sup> (SI eq S1) at 20 °C and with 1.8 ppm atmospheric CH<sub>4</sub>.<sup>30</sup> The same procedure revealed the MDL of  $C_{\rm CO2}$  measurements as 2.39 × 10<sup>-7</sup> mol L<sup>-1</sup>. This MDL is less than the lower range reported in lakes;  $1.00 \times 10^{-5}$  and  $1.30 \times 10^{-5}$  mol L<sup>-1,31,32</sup> This CO<sub>2</sub> MDL is also significantly lower than the dissolved concentration in equilibrium with atmospheric  $CO_2$ , which is 1.54 ×  $10^{-5}$  mol L<sup>-1</sup> (SI eq 1) at 20 °C and with 390 ppm atmospheric  $CO_2$ .<sup>30</sup> We did not test the maximum  $C_w$  that could be measured by the H-ICOS method. According to SI eqs S1 and S2, the volume of headspace injected can be reduced with no theoretical limit in order to avoid the injection of an excessive gas quantity to the UGGA.

The H-ICOS method depends on reaching equilibrium between the water phase and the gaseous headspace in a sampling syringe. The effect of the shaking time on the water/liquid equilibration was determined using water samples prepared in the STR containing known  $C_{CH4}$  and  $C_{CO2}$ . The water samples were gently taken, complemented with CH<sub>4</sub>- and CO<sub>2</sub>-free nitrogen, and vigorously shaken for 0 to 30 s, prior to the headspace injection into the UGGA. SI Figure S2 shows the results obtained, where dissolved gas concentrations are normalized, 1.0 being the final equilibrium concentration. As shown, 10 to 15 s were required to reach equilibrium. According to these results, a shaking time of 20 s was used thereafter as a standard operating procedure.

The H-ICOS method was also tested in the laboratory, under simulated field-conditions; that is, using the prototype and sampling water prepared in the STR, which contained several values of  $C_{\rm w}$ . Figure 2 shows the correlation between the measured and theoretical  $C_{\rm CH4}$  (determined from SI eq 1). A linear response was observed. Similar results were obtained with  $C_{\rm CO2}$  (results not shown). The standard error of the mean (see SI) was estimated to 2.15% for  $C_{\rm CH4}$  and to 1.45% for  $C_{\rm CO2}$  for triplicates.

During the same experiment, the water extracted from the STR was also measured by the M-ICOS method. By comparing the M-ICOS readings with theoretical  $C_{CH4}$ ,  $\alpha$  was determined



**Figure 2.** Measured  $C_{CH4}$  by H-ICOS (white dots) and by M-ICOS (black dots) vs. theoretical  $C_{CH4}$  concentration prepared in a stirred tank reactor. Straight line shows the observed correlations (slope = 1).

to be 8.87 ± 0.30 (eq 2). Figure 2 shows the  $C_{\rm CH4}$  determined by the M-ICOS, according to that  $\alpha$ . For CO<sub>2</sub>,  $\alpha$  was 23.97 ± 1.98 (results no shown). During these M-ICOS determinations, the standard error of the mean was 0.17% for  $C_{\rm CH4}$  and 0.20% for  $C_{\rm CO2}$  for a sampling period of 100 s. Remembering that the parameter  $\alpha$  relates the gas concentration read by the UGGA and the actual  $C_{\rm CH4}$ , this allows the determination of the MDL of the M-ICOS method. Assuming a MDL of the UGGA of 1 ppb for CH<sub>4</sub> and 200 ppb for CO<sub>2</sub>, which correspond to 4.19 × 10<sup>-11</sup> and 8.41 × 10<sup>-9</sup> mol L<sup>-1</sup> at atmospheric pressure and 20 °C, respectively, the M-ICOS method would have a MDL of 2.76 × 10<sup>-10</sup> mol L<sup>-1</sup> for  $C_{\rm CH4}$  and of 1.50 × 10<sup>-7</sup> mol L<sup>-1</sup> for  $C_{\rm CO2}$ . These MDL are similar to those estimated for the H-ICOS method and are lower than the dissolved concentration in equilibrium with atmospheric CH<sub>4</sub> and CO<sub>2</sub>.

To validate the developed response time model (eq 4), several tests were done with different tubing lengths and different gas and liquid flow rates. SI Figure S3 shows an example of the response of the M-ICOS method to sudden change in water concentrations. The model fitted well the experimental data ( $R^2$  for 10 tests = 0.996 ± 0.004), which allowed, after data processing, the determination of  $C_{\rm wm}$  from  $C_{\rm wm}$  data, which is a requirement for online measurements.

3.2. Field-Testing. In the field, the H-ICOS and the M-ICOS methods were tested to assess their operability and also to determine the  $\alpha$  parameter. The H-ICOS method required approximately 3 min per measurement. Triplicate H-ICOS measurements at the same locations and depths of the Lake Guadalupe gave a standard error of the mean of 3.40% for CH<sub>4</sub> and 2.59% for CO2. This error was greater than the error observed during laboratory testing (2.15% and 1.45% for  $C_{\rm CH4}$ and  $C_{CO2}$  respectively), probably because the water samples were independent and taken from slightly different locations because of boat motion. The M-ICOS readings were compared to the  $C_{CH4}$  and  $C_{CO2}$  measurements completed with the H-ICOS method, in order to determine the  $\alpha$  parameter. The  $\alpha$ parameter was measured at several depths of each lake to take into account possible differences in temperature and the concentration of dissolved gas. In different lakes and for CH<sub>4</sub>,  $\alpha$ ranged from 8.05 to 10.73 with a coefficient of variation (see SI) of 9.4%, whereas for CO<sub>2</sub>,  $\alpha$  ranged from 19.17 to 33.11 with a coefficient of variation of 23.5%. These relatively large variations are easily explained by the complexity of  $\alpha$ , which depends on  $Q_{g'}$  K, and H'. In turn, K and H' depend on the temperature of the water and the gas phase, with complex heat transfer between them. However, by comparing  $\alpha$  measured

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within a given lake and maintaining fixed gas and liquid flow rates, the coefficient of variation of  $\alpha$  was reduced to 7.9% for CH<sub>4</sub> and 13.9% for CO<sub>2</sub>. In Lake Guadalupe and Lake Llano, which are relatively deep and were thermally stratified, the correlation between  $\alpha$  and temperature was tested. No trend was observed, probably because of the relatively small temperature gradient (2.7 °C in both cases). However, for the future application of the method in lakes with high thermal stratification, we advise the determination of  $\alpha$  at several depths of the lakes, for potential temperature compensation.

Additionally,  $t_d$  and  $t_r$  were tested in each lake. This was done by rapidly (about 1-2 s) submerging the probe from the surface, where  $C_{CH4}$  was usually low, to a greater depth, where  $C_{CH4}$  was usually higher. SI Figure S4 shows an example of the field response of the M-ICOS method to sudden change in water concentrations in Lake Guadalupe, as well as  $C_{CH4}$ calculated from eq 4. This strategy allowed the field determination of  $t_d$  and  $t_r$ . As expected,  $t_d$  depended on the length of the tubing and the liquid flow rate, while  $t_r$  depended on the gas and liquid flow rates. Little or no effect of tubing length (varying from 6 to 20 m) on  $t_r$  was observed. The concentration profiles vs. time were similar to those observed in the laboratory (SI Figure S3) and fitted well the developed model (eq 4). It was observed that both  $t_d$  and  $t_r$  were stable for fixed gas and liquid flow rates, changing only a few percent over time. From this observation, we decided to determine  $t_r$  and  $t_d$ only every three  $C_w$  profiles. In Lake Goldstream, the diameter of the particulate matter was visually larger than in the other lakes and generated a clogging of the suction probe, i.e. an increase in  $t_{dy}$  during some measurements of the suction probe was observed. In that lake, after each profile, the cleanliness of the probe filter was visually checked, and if dirt had accumulated,  $t_d$  was measured before validating the previous profile, or alternatively, the filter was washed and the profile repeated.

The M-ICOS method was tested to determine  $C_{\rm w}$  profiles. Figure 3 shows an example of the triplicate  $C_{\rm CH4}$  and  $C_{\rm CO2}$  profiles that were obtained sequentially in Lake Guadalupe at the same location. The three determinations gave similar profiles, with lower  $C_{\rm CH4}$  and  $C_{\rm CO2}$  concentrations in superficial water than bottom water, as often observed in lakes.<sup>33–35</sup> A strong  $C_{\rm CH4}$  gradient was observed in Lake Guadalupe between 2.5 and 5 m depth, with an average of  $7.57 \times 10^{-5} \pm 7.10 \times$ 



**Figure 3.** Example of triplicate  $C_{CH4}$  (A) and  $C_{CO2}$  (B) profiles measured in Lake Guadalupe and absolute relative error between triplicate measurements of  $C_{CH4}$  (C).

 $10^{-6}$  mol L<sup>-1</sup> (mean ± sd). Above 2.5 m,  $C_{\rm CH4}$  was lower (1.95 ×  $10^{-6}$  ± 5.43 ×  $10^{-7}$  mol L<sup>-1</sup>) probably due to a combined effect of atmospheric exchange and the presence of DO, which probably promoted CH<sub>4</sub> oxidation (Figure 4 A). A similar  $C_{\rm CO2}$  gradient was observed between 1.0 to 5.0 m depth. By comparing the three profiles, the arithmetic mean of the relative error (relative difference between measurements done at the same depth) was 11.15% for  $C_{\rm CH4}$  and 8.1% for  $C_{\rm CO2}$ . The error was significantly higher when  $C_{\rm w}$  changed abruptly in the ecosystem. For instance, the mean  $C_{\rm CH4}$  error was 56.7% at depths between 2.5 and 3.3 m (Figure 3C), while it was 5.3% outside that depth range. The latter was attributed to error in depth measurements, and a further improvement of the M-ICOS method would involve the coupling of a depth sensor to the suction probe.

Figure 4 shows the depth profiles for  $C_{CH4}$  and  $C_{CO2}$  in the four lakes. In all lakes, except Lake Otto (Figure 4D), clear  $C_{CH4}$  and  $C_{CO2}$  gradients were observed. The absence of  $C_{CH4}$ and  $C_{CO2}$  gradients in Lake Otto can be explained by the fact that Lake Otto is a shallow nonthermokarst lake, classified as oligotrophic,<sup>36</sup> which indicates a low organic carbon input and, therefore, a low methanogenic potential. Additionally, Lake Otto is a lake exposed to constant winds and is therefore well mixed and oxygenated. Contrastingly, steep gradients were observed in Lake Guadalupe (Figure 4A) and in Lake Goldstream (Figure 4C), which receive high carbon input from pollution and thawing permafrost,<sup>37</sup> respectively. It is noticeable that Lake Llano (Figure 4B), which is an unpolluted lake in the same drainage basin as Lake Guadalupe, exhibited moderate gradients. Combined with the DO profiles, also shown in Figure 4, it is clear that when  $C_{CH4}$  and  $C_{CO2}$  gradients were observed, the trend was opposite to that of the DO profile, particularly in the oxycline.

These results confirm that the M-ICOS method allowed the determination of high-resolution C<sub>w</sub> profiles. To the best of our knowledge, this is the first time a method that allows the determination of combined  $C_{CH4}$  and  $C_{CO2}$  data at a frequency of 1 Hz has been reported. In addition, the M-ICOS, allows the determination of the ratio between both parameters, which is an indication of the major processes involved in carbon cycling. CH<sub>4</sub> is mainly produced by strict anaerobic methanogenesis and consumed by methanotrophy, while CO<sub>2</sub> is a product of both anaerobic and aerobic metabolic processes. Thus, the ratio between  $C_{CH4}$  and  $C_{CO2}$  is an indication of aerobic vs. anaerobic organic carbon utilization. Figure 4 shows the  $C_{CH4}/C_{CO2}$  ratios observed in the four lakes. In Lake Guadalupe and Lake Goldstream (Figures 4A and 4C, respectively), a step gradient was observed, just below the oxycline, while in Lake Llano, a moderate gradient was observed; no gradient was observed in Lake Otto. The M-ICOS method is also a convenient tool for the determination of Rq, which describes the predominance of anaerobic over aerobic metabolism.<sup>15</sup> Figure 4 shows the Rq profiles observed in the water column in the four lakes. No Rq gradient was observed in Lake Otto, while in Lake Llano, a moderate gradient was detected. However, in both lakes, the Rq was significantly lower than 1.0, which is indicative of aerobic processes. In Lake Guadalupe, a clear gradient was observed, and Rq values greater than 1.0 were seen below the oxycline, an indicator of the predominance of anaerobic metabolisms. In Lake Goldstream, the Rq was surprisingly high, with values ranging from 4.0 to 25.4; the higher values were observed in the aerobic epilimnion of the lake, contrarily to what is generally reported.<sup>15,26,38,39</sup> It should be pointed out that Lake



Figure 4. Profiles observed in Lake Guadalupe (A), Lake Llano (B), Lake Goldstream (C) and Lake Otto (D) of  $C_{CH4}$  (black dots),  $C_{CO2}$  (black dots), Dissolved oxygen (dotted line), pH (dotted line), CH<sub>4</sub>/CO<sub>2</sub> (black dots), and Respiratory Quotient (dotted line). Note the different *y*- and *x*-axis scales among lakes.

Goldstream has been classified as a dystrophic lake (a brownwater yedoma lake with high dissolved organic carbon concentrations),<sup>40</sup> with high level of  $CO_2$  emissions.<sup>41</sup> In order to obtain high frequency in Rq determinations, a further improvement of the M-ICOS method would require the combination of a high frequency DO sensor and the suction probe.

**3.3. Comments and Recommendations.** The M-ICOS prototype and method, combined with the H-ICOS method for field calibration, allowed the determination of  $C_w$  with a frequency of 1 Hz and with a MDL of  $2.76 \times 10^{-10}$  mol L<sup>-1</sup> for  $C_{CH4}$  and of  $1.50 \times 10^{-7}$  mol L<sup>-1</sup> for  $C_{CO2}$  (see SI Table S2 for specifications). These MDL are significantly lower than the minimum concentrations in lakes, as reported in the literature, and than the  $C_{CH4}$  and  $C_{CO2}$  of freshwater in equilibrium with the atmosphere. The small size of the suction probe (6 mm tubing) and the relatively low liquid flow rate required by the method (600 mL min<sup>-1</sup>) allows for the determination of  $C_w$  with a minimized impact on the ecosystem, including shallow freshwater ecosystems, and with an accurate location and depth awareness.

A possible limitation of the method is the requirement of water sampling, making its application difficult at low air temperature (below freezing), when ice formation may be occurring at the waterline exposed to air. The method application in deep aquatic ecosystems could also be a limitation of the method. In the present work, the M-ICOS method was applied up to a maximum depth of 20 m. Its application in deeper ecosystems would increase  $t_d$ , which is a direct function of the length of the sampling tubing. Additionally, axial dispersion within the water flow in the sampling tubing would also certainly become important,

resulting in a long  $t_r$ . The combined increase in  $t_d$  and  $t_r$ . could be described by a more complex hydraulic residence time model, yielding no theoretical limit to the depth of measurement, despite its practice limitations. The application of the M-ICOS method in oceanographic sciences would present several bottlenecks, compared to the actual state of the art OA-ICOS method.<sup>21</sup> The first one is undoubtedly a very long delay time of several hours between sampling and measurement for depths greater than 1000 m, assuming the same liquid flow speed as observed in the present work. Compared to the 5 min response time reported,<sup>20</sup> these delays are extremely long and may require complex hydraulic flow modeling and an extremey low permeability sampling line. The second bottleneck is the MDL of the M-ICOS method, which is very close to the lower CH<sub>4</sub> concentration range reported in oceans, that is, in the low nmol  $L^{-1}$  range.<sup>7</sup>

Article

Several further developments of the M-ICOS method are recommended. First, to couple the suction probe to a depth sensor and a fast DO sensor, which would allow for a more precise measurement of the  $C_{CH4}$  and  $C_{CO2}$  profiles and a high throughput Rq determination. Second, to use other OA-ICOS detectors to determine the dissolved concentration of analytes such as ammonia, nitrous oxide, oxygen and hydrogen sulfide. OA-ICOS detectors for each of these analytes are now commercially available. And third, a very significant advancement would be to combine our design prototype with stable isotope analyzers, currently available to determine even more specifically the turnover dynamics of different metabolically active and linked analytes.

#### ASSOCIATED CONTENT

#### **Supporting Information**

Extended experimental methods and figures. This material is available free of charge via the Internet at http://pubs.acs.org.

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The manuscript was written through contributions of all authors. All authors have given approval to the final version of the manuscript

#### Notes

The authors declare no competing financial interest

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ecosystems by off-axis integrated cavity output spectroscopy

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Supporting information

17 pages (including references)

Extended description of materials and methods.

Figure S1. (A) Peak response of the UGGA to injections of known CH<sub>4</sub> concentrations;

(B) Integrated area of the peak response to increasing quantities.

Figure S2. Headspace concentrations in the sampling syringe for several shaking times.

Figure S3. Measured CH<sub>4</sub> concentrations by the M-ICOS after sudden changes in  $C_{CH4}$ .

Figure S4. Example of the field response of the M-ICOS method to sudden change in water concentrations and calculated  $C_{CH4}$ .

 Table S1. Main characteristics of the selected lakes.

**Table S2.** M-ICOS prototype specifications.

Material and methods

## UGGA details

Unlike conventional optical methods that rely on low-resolution spectroscopic techniques (e.g., non-dispersive infrared detectors), the UGGA uses a cavity enhanced laser absorption spectrometry technique that quantifies concentrations, based on measurements of high-resolution absorption lineshapes of the target molecules. The fully resolved lineshapes are recorded by tuning independent telecommunication narrow bandwidth (1 MHz) diode lasers, operating near 1600 nm (CO<sub>2</sub>) and 1651 nm (CH<sub>4</sub>, H<sub>2</sub>O), over 1 cm<sup>-1</sup> wide spectral windows, respectively, which straddle the target molecular transitions. Real-time spectroscopic analyses of the measured spectra enables a direct continuous determination of the gas concentrations at rates up to 1 Hz, using Beer's Law, and assessments of gas temperature and pressure in the measurement cell. Because of the use of narrow linewidth tunable lasers to record high resolution, fully resolved lineshapes, no instrument deconvolution is required and cross interferences from other compounds are virtually eliminated.<sup>1</sup>

# Prototype details

The prototype included a continuous flow of  $CH_4$ - and  $CO_2$ -free analytical grade nitrogen (Infra, Mexico or Airgas, U.S.A.) and a continuous flow of water extracted at the desired depth from the freshwater ecosystem through a vacuum line (Figure 1). The gas and the liquid crossed at a gas exchange station. The water line (6 mm internal diameter

polyurethane tubing) included a suction probe to extract continuously the water sample. This probe consisted of a plastic tubing (25 mm diameter, 50 mm long), filled with a washable polyester wool filter to avoid line-clogging by sediment or suspended particulate matter. The probe was fixed on a thin pole for low depth environments or on a 200 g weight for deeper environments to ensure accurate location at the desired sampling depth. The water flow passed through the gas/liquid exchange station, which consisted of a silicone tubing array (Permselect, PDMSXA-1000, Medarray Inc., USA) and then ended in a vacuum graduated glass container with a vacuum gauge. The vacuum glass container was connected to a portable air sampling pump (PCXR4, SKC, USA) that created a fixed vacuum driving force. The water flow rate through the water sampling line was measured volumetrically several times prior to  $C_{\rm w}$  profile determinations. The water flow rate was also controlled through the determination of  $t_d$ . As  $t_d$  is a direct function of the water flow rate,  $t_d$  was a clear indicator of constant water flow rate. It should be noticed that at each sampling location the M-ICOS method was calibrated against the H-ICOS method. As such, a constant water flow rate at approximately 600 mL min<sup>-1</sup> was the main factor to ensure precise  $C_{\rm w}$  measurement.

The gas line (6 mm internal diameter polyurethane tubing) included a 27" gas cylinder containing  $CH_4$ - and  $CO_2$ -free nitrogen (Infra, Mexico or Airgas, U.S.A.). The gas flow rate was regulated at 3 L min<sup>-1</sup> with a mass flow controller (GFC17, Aalborg, USA). After passing through the gas/liquid exchange station, the air was filtered (AcroVent Filter 0.2  $\mu$ m, Pall, USA.) twice to avoid condensed water entering the UGGA detector. The complete set-up weighed approximately 30 kg, was powered by a 50 kWh boat battery, and

was easily operable by a single person from a small portable boat, although a two-person crew allowed an easier operation.

# Details on H-ICOS method

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The H-ICOS, described in the main body of the article, is an adaptation of the traditional gas/liquid equilibration technique, where the equilibrium between a water sample and a CH<sub>4</sub>- and CO<sub>2</sub>-free nitrogen headspace is obtained in a 60-mL sampling syringe. After measuring the syringe's headspace concentration by UGGA, the dissolved gas concentration in the water sample was determined according to Henry's law (eqs S1 and S2), where  $C_w$  is the dissolved gas concentration in the water sample (CH<sub>4</sub> or CO<sub>2</sub>; mol L<sup>-1</sup>),  $C_g^*$  the gas concentration measured in the headspace of the equilibration syringe (mol L<sup>-1</sup>);  $V_1$  and  $V_g$  the water and gas volumes in the syringe, respectively (L); H' the CH<sub>4</sub> and CO<sub>2</sub> air/water partition coefficient (-), defined from eq S2; 1.013 is the conversion factor from atm to bars; R is the universal gas constant (0.082 L atm K<sup>-1</sup> mol<sup>-1</sup>); T is the equilibration temperature (K) at the time of measurement;  $K_H$  is the Henry's law constant at 298.15 K (1.40x 10<sup>-3</sup> and 34.0 x 10<sup>-3</sup> mol L<sup>-1</sup> bar<sup>-1</sup>, for CH<sub>4</sub> and CO<sub>2</sub>, respectively);<sup>2</sup> and  $\beta$  is the temperature dependence coefficient of the Henry's law constant (1700 and 2400 K, for CH<sub>4</sub> and CO<sub>2</sub>, respectively).<sup>2</sup>

$$C_{w} = \frac{\left(C_{g}^{*} \cdot V_{g}\right) + \left(\frac{C_{g}^{*}}{H \cdot} V_{l}\right)}{V_{l}}$$
(S1)

$$H' = \frac{1}{1.013 \cdot R \cdot T \cdot K_H \cdot exp\left[\beta\left(\frac{1}{T} - \frac{1}{298.15}\right)\right]}$$
(S2)

Details on the M-ICOS method

The M-ICOS method consisted of a counter flow of  $CH_4$ - and  $CO_2$ -free nitrogen and water, crossing in the silicone tubing array, as described in the main body of the article. The gas transfer can be described by a diffusion model, according to the Fick's second law (eq 1, repeated here for clarity);

$$\frac{dM}{dt} = 1000 \cdot K \cdot A_M \cdot \left(C_W - \frac{C_g}{H}\right) \tag{1}$$

Since the transferred gas is directed to the gas phase and since the carrier gas contained no CH<sub>4</sub> and CO<sub>2</sub>, eq 1 was modified to eq S3, which was obtained from a simple mass balance, where  $Q_g$  is the gas flow rate (L s<sup>-1</sup>).

$$C_g \cdot Q_g = 1000 \cdot K \cdot A_M \cdot \left(C_w - \frac{C_g}{H'}\right)$$
(S3)

By rearranging eq S3, eq S4 is obtained, in which membrane, gas, and water transfer characteristics ( $Q_g$ , K,  $A_M$  and H) are combined into a single parameter  $\alpha$  for easier calculation.

$$C_w = \frac{C_g \cdot Q_g}{1000 \cdot K \cdot A_M} + \frac{C_g}{H'} = C_g \cdot \left(\frac{Q_g}{1000 \cdot K \cdot A_M} + \frac{1}{H'}\right) = C_g \cdot \alpha$$
(S4)

Equation S4 shows the direct proportionality between  $C_w$  and  $C_g$ , which is directly detected by the UGGA, offering a convenient method for the measurement of the dissolved gas concentration. The parameter  $\alpha$  can be determined by measuring first  $C_w$  by the H-ICOS method and then, with the same sample, by measuring  $C_g$  by the M-ICOS. This can be easily done in the field with actual water samples.

An important issue with the M-ICOS method is the delay time ( $t_d$ ) and the response time ( $t_r$ ) and of the system, as briefly described in the main body of the article. A continuous flow stirred tank reactor (CSTR) model of hydraulic residence time<sup>3</sup> describes well the hydraulic behavior of the system (eq 3).

$$C_{wm} = C_w \cdot \left[ 1 - exp\left( -\frac{t}{t_r} \right) \right] \tag{3}$$

By the derivation of eq 3, eq S5 is obtained;

$$C_w = \frac{dC_{wm}}{dt} \cdot t_r + C_{wm} \tag{S5}$$

Taking into account  $t_d$ , eq 4 was obtained.

$$C_{w,t} = \frac{dC_{wm,t+t_d}}{dt} \cdot t_r + C_{wm,t+t_d}$$
(4)

# Laboratory testing

The precision and linearity of the UGGA were tested by injecting several  $CH_4$  and  $CO_2$  standards (from 2, 5, 20, 50, 200, and 500 ppm High Purity Standards, Infra) and by simultaneously measuring these standards with a gas chromatograph. We used a Clarus-500 (Perkin Elmer, USA) chromatograph equipped with a flame ionization detector (FID) detector and an Elite - Q Plot column for  $CH_4$  and a thermal conductivity detector (TCD) and an Alltech Hayesep D 100/120 column for  $CO_2$ .

The first test of the H-ICOS method was to establish the peak response of the UGGA to several volumes and  $CH_4$  and  $CO_2$  concentrations injected in the gas line. With this purpose, a continuous CH<sub>4</sub>- and CO<sub>2</sub>-free nitrogen gas flow rate of 3 L min<sup>-1</sup> was established and controlled with a mass flow controller; then, 0.1 to 40 mL of 2 to 500 ppm CH<sub>4</sub> or 20 to 500 ppm CO<sub>2</sub> standards were injected. The peaks obtained were integrated (concentration over time using Wolfram Mathematica 8.0, USA). Then, synthetic water samples, containing a known  $C_w$ , were prepared in a 3 L lab-scale STR (ez-Control, Applikon, Netherlands) by injecting a continuous flow of 2 to 500 ppm standard gases in 2.5 L tap water with strong mixing (800 rpm), until saturation was obtained. To establish the time required to reach saturation in the STR, prior experiments were conducted by injecting air or nitrogen and measuring the dissolved oxygen concentration until 100% or 0% saturation was obtained (HI2400, Hanna Instruments, USA). Water containing a known dissolved gas concentration was prepared by mixing and gassing times at least two times greater. The dissolved gas concentration in water, with all standards, was established according to Henry's law (eqs S1 and S2). Samples of water were taken with a disposable

60 mL syringe, according to the H-ICOS method, and several shaking times were tested to establish the time required to reach equilibrium between the water sample and the headspace of the syringe. The headspace was injected in the UGGA and also measured by gas chromatography. Then, the M-ICOS method was tested, using the same water samples, in order to establish  $\alpha$  (eq 2). Finally,  $t_d$  and  $t_r$  were determined with the M-ICOS method, by switching between water containing CH<sub>4</sub> and CO<sub>2</sub> to degased water using a 3-way valve. These experiments were also used to check the  $t_r$  model developed (eq 4).

# Field-testing

The prototype and both methods were tested in four different lakes with contrasting climates, ecologies, and morphologies. The first lake was a eutrophic subtropical reservoir located in the Mexico metropolitan area (Lake Guadalupe, 19.6310 N, 99.2567 W). The second lake was a mesotrophic subtropical reservoir located in the same drainage basin as Lake Guadalupe (Lake Llano, 19.6577 N, 99.5069 W). Both subtropical lakes have been previously described.<sup>4</sup> The third lake was a shallow yedoma-type, thermokarst lake (Lake Goldstream, 64.9156 N, 147.8486 W), and the fourth lake was a shallow non-thermokarst lake (Lake Otto, 63.8413 N, 149.0384 W). Lake Goldstream has been previously described.<sup>5-6</sup> Lake Otto is a shallow tundra lake subject to high winds, which has been hitherto studied.<sup>7</sup>

# Statistical and error analysis

The method detection limit (MDL)<sup>8</sup> of both H-ICOS and M-ICOS were determined as the minimum concentration that can be distinguished from background noise with 99% confidence. The goodness of the correlation between the experimental data and the

response time model (eq 4) was quantified with the coefficient of determination (R<sup>2</sup>). Measurement error was determined by the coefficient of variation, defined as the standard deviation divided by the arithmetic mean, and by the standard error of the mean, defined as the coefficient of variation divided by the square root of the replicate number. We also measured the signal to noise ratio of the UGGA, which is the arithmetic mean of the UGGA reading, divided by the standard deviation. Accuracy was calculated as the absolute difference between the measured and the expected concentration relative to expected concentration. Dynamic range was calculated as the logarithmic ratio between the maximum and the minimum concentrations measured. Maximum concentration was theoretically estimated from the UGGA specifications.



**Figure S1.** (A) Example of peak response of the UGGA to triplicate injection of 5 mL nitrogen containing 2 ppm of  $CH_4$  and the minimum injected  $CH_4$  quantity that was distinguishable from the background (inner Figure); (B) Integrated area of the peak response to increasing  $CH_4$  (white dots) and  $CO_2$  (black dots) quantities.



Figure S2. Normalized  $CH_4$  (white dots) and)  $CO_2$  (black dots) headspace concentrations in the sampling syringe for several shaking times. Error bars indicate the standard deviation of the triplicate measurements.



**Figure S3.** Example of measured  $CH_4$  concentrations by the M-ICOS after sudden changes in  $C_{CH4}$ ; decreasing concentration gradient (white dots) and increasing concentration gradient (black dots); normalized concentration, 1.0 being the initial or final steady state concentration.



**Figure S4.** Example of the field response of the M-ICOS method to sudden change in water concentrations (black dots), as well as  $C_{CH4}$  calculated from eq 4 (white dots); normalized concentration, 1.0 being the final steady state concentration.

Lake	Guadalupe	Llano	Goldstream	Otto
Area (km <sup>2</sup> )	4.5	0.06	0.1	0.51
Mean depth (m)	13.3	9.49	3.3	2.5
$TSI^{a}(-)$	Hypereutrophic	Mesotrophic	Distrophic	Oligotrophic
Secchi Depth (m)	0.55	2.13	1.0	1.6
pH	7.2	6.77	7.78	7.68
Surface DO (mg $L^{-1}$ )	2.89	6.48	8.50	7.94
Bottom DO (mg $L^{-1}$ )	0.00	4.02	0.01	7.93
Surface temperature (°C)	21.75	13.88	18.18	12.46
Bottom temperature (°C)	17.35	11.62	11.62	12.28
Mixed layer <sup>b</sup> (m)	0.50	0.50	1.10	2.5
Temperature gradient in	2.80	2.44	3.01	No
the thermocline <sup>c</sup> ( $^{\circ}C m^{-1}$ )				thermocline

Table S1. Main characteristics of the selected lakes.

<sup>a</sup> Trophic State Index, measured as given by Carlson.<sup>9</sup>

<sup>b</sup> Layer from the surface of the water to the depth at which temperature declined at a rate lower than 1 °C m<sup>-1</sup>.

<sup>c</sup> The thermocline was considered to be the layer between the depth at which temperature started to decline at a rate higher than 1 °C m<sup>-1</sup>, to the depth at which it started to decline at a rate lower than 1 °C m<sup>-1</sup>.

 Table S2. M-ICOS prototype specifications.

Weight (complete prototype)	30 kg			
Shipping dimension (suitcase)	$0.47 \times 0.36 \times 0.18$ m (UGGA)			
Power consumption (external battery)	12 VDC, 5.8 A (70 W)			
Operating temperature	0 + 40 °C			
Startup time	2 min			
Response time UGGA	8 s			
Response time M-ICOS	$9.77 \pm 1.01 \text{ s}$			
Calibration time <sup>a</sup>	15 min			
Data acquisition frequency	1 Hz			
Signal to noise ratio CH <sub>4</sub>	$1520 \pm 415$			
Signal to noise ratio CO <sub>2</sub>	$1803 \pm 344$			
Standard error of the mean $C_{CH4}$ determination (n = 100) <sup>b</sup>	0.17 %			
Standard error of the mean $C_{CO2}$ determination (n = 100) <sup>b</sup>	0.20 %			
Method detection limit $C_{CH4}$	$2.76 \ge 10^{-10} \mod L^{-1}$			
Method detection limit $C_{CO2}$	$1.50 \ge 10^{-7} \mod L^{-1}$			
Accuracy for CH <sub>4</sub>	10.64%			
Accuracy for CO <sub>2</sub>	12.22%			
Dynamic range for CH <sub>4</sub>	5.00			
Dynamic range for $CO_2$	5.18			
<sup>a</sup> Triplicate calibration with <b>H</b> ICOS method and determination of t and t				

<sup>a</sup> Triplicate calibration with H-ICOS method and determination of  $t_r$  and  $t_d$ .

<sup>b</sup> n = number of replicates

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